

Distribution of Copper, Cadmium, and Lead in Soils from Former Industrialized Urban Areas of Beijing, China

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Abstract Concentrations and distributions of cadmium, copper, and lead in soils from several industrialized urban areas of Beijing, China were investigated. The mean concentration of Cd in surface soils was not significantly greater than the regional background concentration, while those of Cu and Pb were significantly greater than regional background concentrations. All soil profiles exhibited a decreasing trend in concentrations of Cd, Cu, and Pb with depth. Concentrations of Cu and Pb were greatest at a depth of 0–80 cm, and exceeded the regional concentrations. Concentrations of Cd, Cu, and Pb at some industrial sites would necessitate active remedial actions.

Keywords Heavy metals · Soil contamination · Industrialized urban areas

Due to increased demand for housing in China, industrial complexes are being moved from urban areas to the suburbs and the former industrial districts are being developed as housing. The people living in communities built on former industrial sites may be at risk of cancer and other adverse health effects, even long after the source is removed (Stephens et al. 2004). Thus, contamination of former industrial sites is a critical and pressing environmental issue in China. Some studies have been undertaken on metal contamination of soils as a result of industrial activities, especially those that took place in cities (Weiss et al. 1994; Nriagu and Pacyna 1988). Soils at former industrial sites are particularly prone to having relatively great concentrations of Cu, Cd, and Pb that are derived from the discharge of a variety of industrial pollutants in the form of gases, liquids, and solids to land (Fakoyade and Onianwa 2002). Excessive Pb, Cd, and Cu exposure can cause severe health effects in humans (Flora 2002; Nordberg 2003; Yaman and Akdeniz 2004).

The industrialization of Beijing began in the 1950s. The number of industrial plants reached approximately 20,000 by the end of the 1990s. With Beijing's successful bid to host the Olympic Games in 2008 and the rapid expansion of urban areas in recent years, most industrial plants in the city are being shut down or removed from downtown to sub-cities and the former industrial sites have been converted into commercial and residential properties. However, there have been relatively few such studies conducted in these industrial areas inside Beijing City. For the first time, we have investigated total copper, cadmium

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and lead concentrations in soils, interpreted the profiles of their concentration versus depth, and discussed their contaminations in former industrial sites of Beijing, China. The results obtained in our studies can be used by authorities to identify former sources of pollution, locate polluted areas for remediation and to choose the proper remediation strategy.

Materials and Methods

Five representative industrial sites (denoted A, B, C, D and E) in the industrialized area located approximately 1–8 km southeast of downtown Beijing, China were studied. The description of the five former industrial plants corresponding to the sites is presented (Table 1). Sampling points were selected depending on the location of production and waste treatment sectors, which took up most of the site and were potentially important sources of contamination. Fifty-eight surface soil sample points and 27 profile pits were selected (Fig. 1). Four sub-samples of the topsoil (0–10 cm) were collected at each sampling point and mixed thoroughly to get a representative surface soil sample. The deep layer soil samples were taken from each profile pit at depths of 20, 80, 180, and 400 cm. Therefore, a total of 58 representative surface soil samples and 108 deep layer soil samples were obtained. Prior to analysis, stones and foreign objects were removed by hand, and the soil samples were air-dried at room temperature for 7 day, gently crushed in an agate mortar, completely passed through a nylon sieve of 100 meshes (0.149 mm), and then stored in glass bottles. Aliquants of approximately 0.250 g sample were accurately weighed in a 250 mL Pyrex Erlenmeyer flask and 10 mL of 8.5 M HNO₃ were added. The solution was heated on a hot plate to ~95°C without boiling and this temperature was maintained for 15 min. After cooling to less than 70°C, 5 mL of 17 M HNO₃ were added and the sample was refluxed for 30 min at ~95°C without boiling. This step was repeated a second time. The sample was evaporated to ~5 mL without boiling. After cooling to less than 70°C, 2 mL of water were added

followed by the slow addition of 3 mL of 8.8 M H₂O₂. The solution was then heated until effervescence subsided. After cooling to less than 70°C 10 mL of 12 M HCl were added and the sample was refluxed for 15 min. without boiling. After cooling to room temperature, the sample was filtered and diluted to 25 mL with water. The full description of the method is reported in the related USEPA standard 3050B (1996). Total concentrations of Cu, Cd, and Pb in the digests were analyzed using inductively coupled plasma-mass spectrometry (ICP-MS) (7500a Agilent Technologies). Standard reference materials, GBW07427 soils, obtained from Center of National Standard Reference Material of China, were inserted for quality assurance and quality control (QA/QC) procedures. Average recoveries of Cu, Cd, and Pb in triplicate analysis were 96 ± 2.1%, 99 ± 1.4%, 104 ± 5.0%, respectively. The detection limits were 0.04 mg/kg for Cu, 0.01 mg/kg for Cd and 0.01 mg/kg for Pb.

Statistical analyses were conducted by use of Microsoft Excel and SPSS 10.01 (SPSS Inc., USA). The distributions of concentrations were tested to determine if they approximated the normal probability function with the Kolmogorov–Smirnov method. Concentrations of Cu, Cd, and Pb from each site were analyzed by use of a one way Analysis of Variance (ANOVA) to determine if they were different from one another. Differences between means were tested by Turkey's means comparison test for Post Hoc Multiple Comparisons. A significance level of $p < 0.05$ was used throughout the study.

Results and Discussion

Summary statistics for Cu, Cd, and Pb in the surface soil samples at five industrial sites are presented (Table 2). Concentrations of Cd in surface soils from Site A ranged from 0.14 to 9.2 mg/kg, dw with a mean concentration of 1.9 mg/kg, dw, which was significantly greater than any other Site ($p < 0.05$). The standard deviation (3.6 mg/kg, dw) of Cd concentrations in soils of Site A was the greatest of all the Sites. This result suggests strong variability due to

Table 1 Descriptions of five industrial plants in urban areas of Beijing, China

Plant	Area (10 ⁶ m ²)	Year of establishment	Produced products	Environmental accident events	Proposed land use
A	64.94	1958	Chemical agents, fluorescent and photosensitive materials	19	Residential area
B	7.95	1956	Pesticides, PVC plastic and rubber products	3	Residential area
C	135	1959	Coke, coal gas, ammonium-sulfate, benzene, naphthalene and hydroxybenzene	6	Real estate
D	18.5	1954	Pesticides, Lacquer, architectural and industrial pigments	2	Real estate
E	40	1956	Dyestuffs and painting products and vitriol	0	Real estate

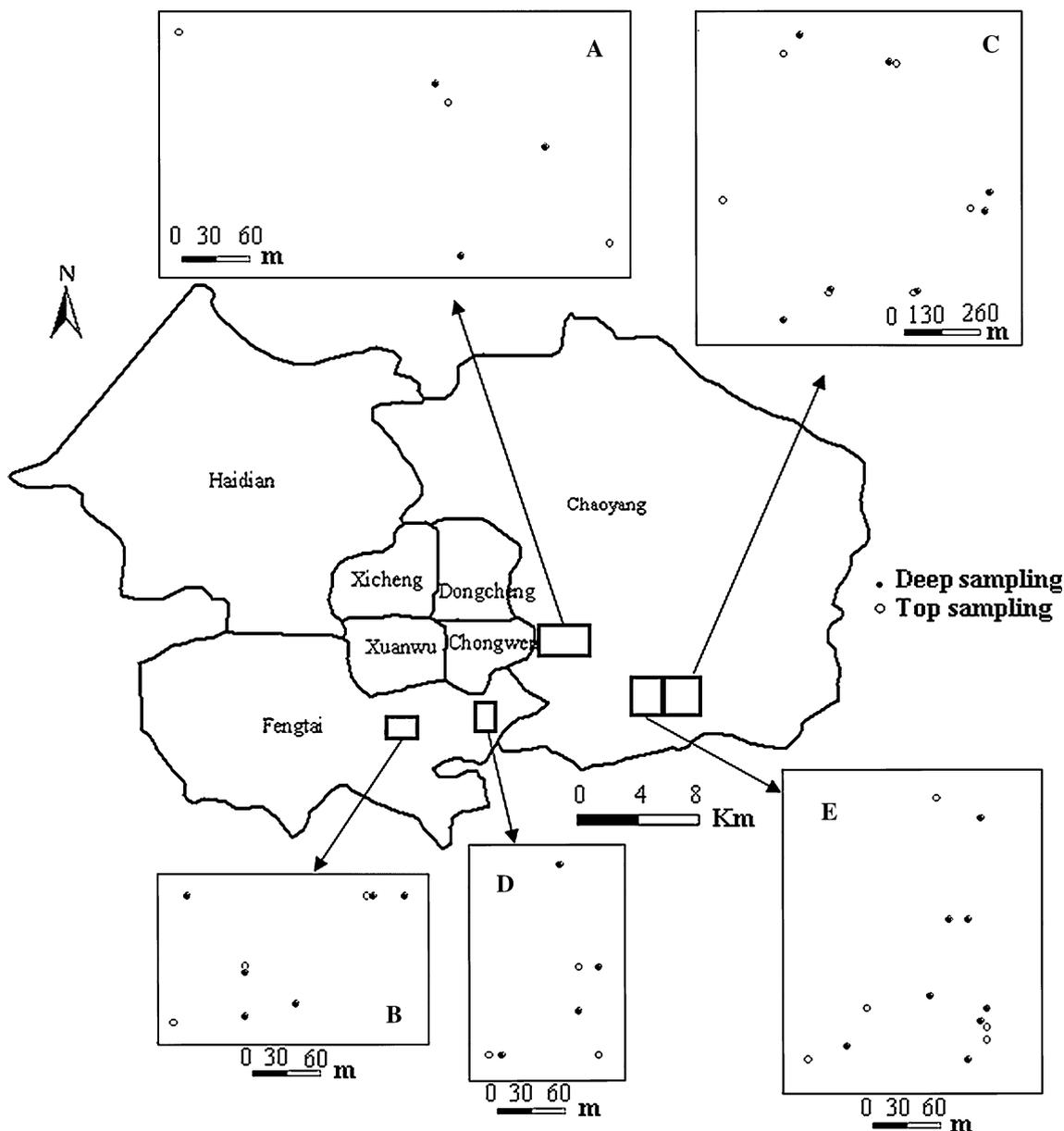


Fig. 1 Sampling scheme at the industrialized urban area of Beijing, China

the presence of different anthropogenic sources of Cd at Site A. The relatively great concentrations of Cd in soils of Site A might be attributed to metal-containing agents (including those containing cadmium) that were produced in the 1960s with relatively primitive and inefficient production techniques without any environmental treatment. However, significant differences in Cd concentrations among the other Sites (namely B, C, D, and E) were not observed, because the other Plants (with the exception of Plant A) did not produce any cadmium-containing products. This may explain the relatively lesser concentrations of Cd at the other Sites. Generally, the mean concentration of Cd in surface soils at industrial sites (0.40 mg/kg, dw) was not

significantly greater (one-sample T-test; $p > 0.05$) than the mean concentration of Cd in less industrialized reference areas that represented the general background concentration of soils in the region (0.15 mg/kg, dw). Hence, the surface soils from the Beijing industrial sites were not significantly contaminated by cadmium. Soils at Site E had a greater mean concentration of Cu (92 ± 78 mg/kg, dw) than those from Sites B, C, and D ($p < 0.05$) (Table 2). Concentrations of Cu in the soils from Site E varied greatly based on range ($14\text{--}2.2 \times 10^2$ mg/kg, dw) and standard deviation (78 mg/kg, dw). This could be attributed to the fact that Site E produced dyestuffs, painting products and vitriol, which required a substantial amount of Cu-contained materials

Table 2 Descriptive basic statistics of concentrations of Cd, Cu, and Pb (mg/kg, dw) in the surface soils at different industrial sites

Site	Mean*	Median	SD	Min	Mx
Cadmium					
A (n = 6)	1.9 b	0.46	3.6	0.14	9.2
B (n = 14)	0.24 a	0.21	0.13	0.10	0.58
C (n = 14)	0.22 a	0.16	0.20	0.09	0.88
D (n = 10)	0.27 a	0.24	0.13	0.13	0.54
E (n = 14)	0.18 a	0.14	0.10	0.08	0.43
Overall (n = 58)	0.40	0.18	1.2	0.08	9.2
Non-polluted soils (n = 117) ^a	0.15	0.11	0.11	0.03	0.63
Copper					
A (n = 6)	40 ab	34	16	24	61
B (n = 14)	29 a	26	15	14	70
C (n = 14)	32 a	23	18	14	69
D (n = 10)	37 a	33	16	21	75
E (n = 14)	92 b	55	78	14	2.2×10^2
Overall (n = 58)	48	33	47	14	2.2×10^2
Non-polluted soils (n = 117) ^a	20	19	6.3	6.0	38
Lead					
A (n = 6)	52 a	48	25	22	92
B (n = 14)	1.1×10^2 a	52	1.9×10^2	37	7.7×10^2
C (n = 14)	83 a	56	73	22	3.0×10^2
D (n = 10)	1.1×10^2 a	99	61	34	2.4×10^2
E (n = 14)	46 a	44	24	17	1.1×10^2
Overall (n = 58)	82	54	1.1×10^2	17	7.7×10^2
Non-polluted soils (n = 101) ^a	25	25	5.1	12	38

* For the indicated elements, those values not showing a common small letters (a, b) are significantly different at the level ($p < 0.05$)

^a Chen et al. (2004)

such as $C_{22}H_{16}N_8Cu$, $C_{32}H_2Cl_{14}Cu$, and $CuCl$. There was no significant difference in the concentrations of Cu among Sites B, C, and D ($p > 0.05$). The mean concentration of Cu in surface soils from all of the industrial sites (48 mg/kg, dw) was significantly greater than that for the reference soils (20 mg/kg, dw). Therefore, the industrial soils at these sites contained elevated levels of Cu. Relatively greater mean concentrations of Pb ($1.1 \times 10^2 \pm 1.9 \times 10^2$ mg/kg, dw and $1.1 \times 10^2 \pm 61$ mg/kg, dw) were observed in surface soils from Sites B and D, although there was no statistically significant difference in the concentration of Pb between industrial sites. Pesticides, PVC plastic and rubber products were produced at Site B. In addition Pb-containing powder was used for production of industrial and architectural pigments at Site D, which might have caused

greater concentrations of Pb in soils at Sites B and D. The mean Pb concentration in all the industrial soil samples was significantly greater than that of soils in the reference locations. Thus, we concluded that the surface soils collected from Beijing industrial sites were contaminated by Pb.

Concentrations of Cd (Mean \pm SD) at different depths of soil profiles taken from industrial sites are presented (Table 3). Mean concentrations of Cd in soil profiles tended to decrease with increasing depth. Cd concentrations at depths of 20–80 cm in soils from Site A were significantly greater than those at depths of greater than 80 cm. This result indicated severe heavy metal pollution at the layers of 20–80 cm. However, the Cd concentrations in soils deeper than 80 cm from Site A tended to be consistent with

Table 3 Cd concentrations (mg/kg, dw) in the soil profiles taken from different industrial sites

Site	20 cm	80 cm	180 cm	400 cm
A (n = 3)	3.4 ± 5.1	1.0 ± 1.4	0.13 ± 0.080	0.15 ± 0.070
B (n = 5)	0.16 ± 0.080	0.11 ± 0.020	0.10 ± 0.030	0.12 ± 0.020
C (n = 7)	0.23 ± 0.29	0.17 ± 0.17	0.17 ± 0.20	0.10 ± 0.020
D (n = 4)	0.21 ± 0.060	0.15 ± 0.090	0.10 ± 0.040	0.10 ± 0.020
E (n = 8)	0.15 ± 0.080	0.17 ± 0.19	0.12 ± 0.050	0.11 ± 0.030
Overall (n = 27)	0.54 ± 1.75	0.25 ± 0.50	0.12 ± 0.10	0.11 ± 0.030

the mean concentration in non-polluted soils (0.15 mg/kg, dw). This suggests that Cd pollution is not an issue below 80 cm and that leaching from less deep soils to deeper soils has been minimal. There was no statistically significant difference in concentrations of Cd among the depths of soil profiles at the other industrial sites, suggesting there was no significant Cd pollution in the deep layers of the other sites. Concentrations of Cu in the soil profiles decreased with increasing depth (Table 4). Mean concentrations of Cu at depths of 20 and 80 cm at Sites A, B and C were significantly greater than not only those at the depths of 180 and 400 cm, but also the mean concentration in reference (20 mg/kg, dw). These results demonstrate that significant amount of Cu at Sites A, B, and C leached to at least 80 cm, although there was no significant contamination at the depth of 180 cm. Copper in soils from Sites D and E has penetrated to at least 180 cm because its concentrations in soils at the depths of 180 cm were significantly greater than the concentration of Cu in reference soils (20 mg/kg, dw). Thus, the soils at depths of 0–180 cm at Sites D and E were also contaminated by Cu. Mean concentrations of Cu at the depths of 20–80 cm in the soil profiles were significantly greater than not only those at the depths of 180 and 400 cm, but also reference soil. It could be concluded that the soils at the depths of 0–80 cm at Beijing industrial sites were contaminated by Copper. Concentrations of Pb in industrial soils from all sites decreased with increasing depth of the profiles (Table 5). Mean concentrations of Pb

at depths of 20, 80, and 180 cm at Site D were significantly ($p < 0.05$) greater than those at 400 cm (26 mg/kg), similar to the concentration in reference soils (25 mg/kg). Mean concentrations of Pb at 20 and 80 cm at Sites B and C were significantly greater than those at the depths of 180 and 400 cm. However, concentrations of Pb among the depths of the profiles at the Site E were not statistically significant. Generally, concentrations of Pb at depths of 20–80 cm at the industrial sites were significantly greater than not only those at depths of 180 and 400 cm, but also those of regional reference soils (25 mg/kg, dw). Hence, it can be concluded that soils at 0–80 cm for the industrial sites have been contaminated by Pb. Furthermore, mean concentrations of Pb at a depth of 400 cm in soil profiles at Sites B and D were greater than the regional reference soils. Therefore, Pb has leached at least 400 cm downward at these industrial sites.

Most countries have promulgated legislation on soil protection. Soil clean-up standards (SCSs) are often used as trigger criteria to determine the need for an in-depth soil investigation or for remediation. The needs for further soil investigation or remedial action are related to the exceedance of an SCS. However, there is no soil clean-up standard for different land uses for China to date. In this study, we used SCSs for Cd, Cu, and Pb for residential land use for some other countries to determine if there is a need to conduct an in-depth soil investigation, or to perform remedial action (Table 6). Mean concentrations of Cd in

Table 4 Cu concentrations (mg/kg, dw) in the soil profiles taken from different industrial sites

Site	20 cm	80 cm	180 cm	400 cm
A (n = 3)	38 ± 20	22 ± 7.8	13 ± 6.6	9.4 ± 7.0
B (n = 5)	32 ± 20	24 ± 10	14 ± 8.8	17 ± 7.4
C (n = 7)	26 ± 20	22 ± 6.6	13 ± 8.3	14 ± 4.0
D (n = 4)	40 ± 8.1	37 ± 13	30 ± 23	16 ± 3.3
E (n = 8)	85 ± 80	45 ± 31	37 ± 27	17 ± 6.7
Overall (n = 27)	47 ± 51	30 ± 20	23 ± 18	15 ± 5.7

Table 5 Pb concentrations (mg/kg, dw) in the soil profiles taken from different industrial sites

Site	20 cm	80 cm	180 cm	400 cm
A (n = 3)	47 ± 15	26 ± 17	19 ± 5.6	3.4 ± 1.8
B (n = 5)	78 ± 47	84 ± 67	30 ± 21	28 ± 22
C (n = 7)	67 ± 48	59 ± 68	27 ± 24	19 ± 5.8
D (n = 4)	1.0 × 10 ² ± 41	99 ± 99	68 ± 77	26 ± 17
E (n = 8)	42 ± 16	41 ± 20	34 ± 19	17 ± 4.0
Overall (n = 27)	66 ± 40	59 ± 57	34 ± 32	19 ± 12

Table 6 Soil clean-up standards for residential landuse for some countries (Provoost et al. 2006)

Country	Cadmium (mg/kg, dw)	Copper (mg/kg, dw)	Lead (mg/kg, dw)
Belgium ^a	6.0	4.0 × 10 ²	7.0 × 10 ²
Netherlands	12	1.9 × 10 ²	5.3 × 10 ²
Germany ^b	20	NA	4.0 × 10 ²
France	20	1.9 × 10 ²	0 × 10 ²
Sweden	0.40	1.0 × 10 ²	80
Norway	3.0	1.0 × 10 ²	60
Canada	10	63	1.4 × 10 ²
Switzerland	20	1.0 × 10 ³ ^c	1.0 × 10 ³ ^d
USA	37	3.1 × 10 ³	4.0 × 10 ²

NA not applicable

^a Soil degree named Vlarebo from July 8, 2002

^b Standards applicable as national legislation for 'wirkungspad Boden-Mensch' (exposure path soil–humans)

^c 1000/4 related to the soil clean-up standard as total concentration and soluble concentration. The clean-up standard of 1,000 mg/kg dm was used in this comparison

^d 1000/0.1 related to the soil clean-up standard as total concentration and soluble concentration. The clean-up standard of 1,000 mg/kg dm was used in this comparison

the soils at all Sites, with the exception of Site A were substantially less than the SCSs for most countries. However, the maximum values of Cd in surface soils samples from all the Sites (Table 2) did exceed the SCSs for Sweden (0.40 mg/kg). However, these concentrations were less than 0.40 mg/kg, dw at greater depths of soil (Table 3). Site A had mean concentrations of Cd at depths of 0–80 cm greater than the SCS of Sweden (0.40 mg/kg), and some were even greater than the SCSs of Norway and Belgium, which are 3.0 and 6.0 mg/kg, dw, respectively. So, remedial action or at least proper management for Cd contamination in the soils at the depths of 0–80 cm at Site A is necessary. Based on the SCSs of Canada (63 mg/kg), the soils at depths of 0–20 cm at Site E should be remediated. Alternatively, no remediation would be needed at the other industrial sites because the mean concentrations of Cu in the soils were less than the SCSs for all the countries (Table 6). Concentrations of lead in soils at 20 cm for Sites A and C, 0–80 cm for Site B, and 0–180 cm for Site D exceed the SCS of lead for Norway (60 mg/kg). Only concentrations of Pb in the soils at depths of 0–10 cm for Sites A, C, and D, 0–80 cm for Sites B and D exceed the SCS of Pb for Sweden (80 mg/kg). None of the concentrations of Pb exceeded the SCSs for any other countries (greater than 1.4×10^2 mg/kg). Based on our findings, some soils at industrial sites will require in-depth soil investigation or in some cases remediation. In conclusion, remediation action or proper management for Cd, Cu, and Pb contamination is necessary at some industrial sites which would be used for residential areas, based on the soil clean-up standards for some countries.

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